



HAL
open science

IMPACT OF WATER RADIOLYSIS ON URANIUM DIOXIDE CORROSION

A. Traboulsi, Johan Vandendorre, G. Blain, J. Barbet, M. Fattahi

► **To cite this version:**

A. Traboulsi, Johan Vandendorre, G. Blain, J. Barbet, M. Fattahi. IMPACT OF WATER RADIOLYSIS ON URANIUM DIOXIDE CORROSION. 7th EC FP CP FIRST-Nuclides, 2nd Annual Workshop 2013, Nov 2013, Antwerpen, Belgium. 2013. in2p3-00911344

HAL Id: in2p3-00911344

<https://in2p3.hal.science/in2p3-00911344v1>

Submitted on 29 Nov 2013

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

IMPACT OF WATER RADIOLYSIS ON URANIUM DIOXIDE CORROSION

Johan Vandendorre^{1*}, Ali Traboulsi¹, Guillaume Blain¹, Jacques Barbet², Massoud Fattahi¹

¹ SUBATECH, UMR 6457, Ecole des Mines de Nantes, CNRS/IN2P3, Université de Nantes,
4 rue Alfred Kastler, BP 20722, 44307 Nantes cedex 03, France.

² Cyclotron ARRONAX, 1 Rue ARRONAX, 44817 St-Herblain Cedex France

* Corresponding author: johan.vandendorre@subatech.in2p3.fr.

Abstract

In this work, the oxidation of UO_2 by the radiolysis products of water at the solid/solution interface is investigated in function of the dose under open and close atmospheres. Irradiation is realized by He^{2+} beam provided by the ARRONAX cyclotron with an energy of 66.5 MeV and a dose rate of 4.37 kGy/min. The aim of this investigation is to determine the effect of the atmosphere (presence of H_2) and the dose on the UO_2 oxidation in order to couple for the first time (1) characterization of the secondary oxidized phases, (2) quantification of H_2O_2 and H_2 produced by water radiolysis and (3) determination of the quantity of uranium released into the solution. The kinetic of the solid surface oxidation is followed by Raman spectroscopy. H_2O_2 and H_2 are measured respectively by UV-VIS spectrophotometry and micro Gas Chromatography ($\mu\text{-GC}$). Inductively Coupled Plasma Mass Spectrometry (ICP-MS) is used to quantify the soluble uranium species released into the solution. Our results show that He^{2+} irradiation of water induced oxidation of the UO_2 surface which depended on the atmosphere and the dose rate. We present below the results obtained for two samples irradiated at a dose of 8.73 kGy under the two atmospheres mentioned above. Results shows the Raman spectra of unirradiated UO_2 (a) and a sample irradiated under open atmosphere (b, c, d). The spectra of the irradiated sample are obtained 0.3, 42 and 169 hours after contact with the irradiated water. When comparing the two spectra, 2 identical Raman signals can be observed at 445 and 560 cm^{-1} . The first one characterizes the U-O stretch in the fluorite structure of UO_2 and the second one indicates the presence of defects in the matrix. After irradiation, two Raman signals appear at 820 and 865 cm^{-1} . These vibration bands indicate the presence of studtite ($\text{UO}_4 \cdot 4(\text{H}_2\text{O})$) formed by oxidation of the UO_2 surface by H_2O_2 produced by water radiolysis. We have verified that studtite is not formed by oxidation of the UO_2 surface by O_2 present in the atmosphere. For the irradiated sample, $G(\text{H}_2\text{O}_2)$ formed by water radiolysis is $0.06\text{ }\mu\text{mol/J}$ and the quantity of uranium species released in the irradiated solution is $11.5 \cdot 10^{-7}\text{ mol/L}$. For the sample irradiated under close atmosphere, oxidation is much slower due to the inhibition effect of H_2 produced by water radiolysis. In these conditions, $G(\text{H}_2\text{O}_2)$ and $G(\text{H}_2)$ are respectively 0.1 and $0.02\text{ }\mu\text{mol/J}$. The concentration of Uranium species in the solution is $5.3 \cdot 10^{-7}\text{ mol/L}$ which is comparable to that published in the literature.

$G(\text{H}_2\text{O}_2)$ obtained in open atmosphere is two fold lower than that obtained in close one. In this last condition, H_2O_2 yield is equal to that of ultra-pure water which means that this molecule was not consumed. Moreover, $G(\text{H}_2)$ obtained in this work is lower than that obtained by radiolysis of ultra-pure water in similar conditions. From this, we can conclude

that H₂ has an inhibition effect on the UO₂ oxidation and this inhibition does not take place by direct effect on H₂O₂ but by interaction between H₂ and the UO₂ surface.

Introduction

This paper deals with the radiolytic corrosion at the UO₂ surface. We study the impact of water radiolysis onto the corrosion of the grain boundaries (GB) present at the TRISO particle surface. The influence of H₂ produced by water radiolysis on the UO₂ corrosion is also studied. More details on the context of this study is published elsewhere (Kienzler et al., 2012). The aim of this work was to investigate experimentally the effect of molecular radiolytic species (H₂ and H₂O₂) produced by localized ⁴He²⁺ radiolysis of water on UO₂ corrosion in function of different parameters. Water in contact with UO₂ particles was then irradiated by ⁴He²⁺ beam at different absorbed doses and under different atmospheres in order to evaluate the effect of H₂ and the absorbed dose on the corrosion process. To fulfill this work, we chose to investigate the corrosion of the surface of solid particles by Raman Spectroscopy which is known to be very efficient in surface characterization of solid state materials. Soluble uranium species leached into the solution were analyzed by Inducted Coupled Plasma Mass Spectrometry (ICP-MS) technique. H₂O₂ and H₂ produced by water radiolysis were respectively quantified by UV-VIS Spectrophotometry and Gas Chromatography (μ-GC) in order to calculate their radiolytic yields, compare them to those of pure water and clarify the role of these two species in the UO₂ corrosion.

Material and Methods :

The experimental section (analytical tools and irradiation conditions) is described elsewhere (Kienzler et al., 2012). 3 irradiations were performed for 3 sets of sample: (1) irradiation under open to air atmosphere (oxidative conditions without H₂), (2) close with air atmosphere (oxidative conditions with H₂ produced by water radiolysis), and (3) close with Ar/H₂ (4%) atmosphere (reductive atmosphere where H₂ is initially added into the system before irradiation by the Ar/H₂ (4%) gas).

Results

1. UO₂ Surface Characterization

UO₂ TRISO particle is constituted of a UO₂ based kernel (500 μm of diameter) composed by grains and grain boundaries (**Figure 1**). The grain size has been measured between 10 and 20 μm. After irradiation, SEM pictures were realized on the surface particles in order to determine the zones most affected by the corrosion process (**Figure 2**). **Figure 2** clearly shows that regardless the irradiation atmosphere, the grain boundaries are a lot more degraded than the UO₂ grains. It appears too that the corrosion did not occur homogeneously on the entire surface. Some regions of the grains surface are indeed more corroded. The corrosion can be considered as localized to the GB.

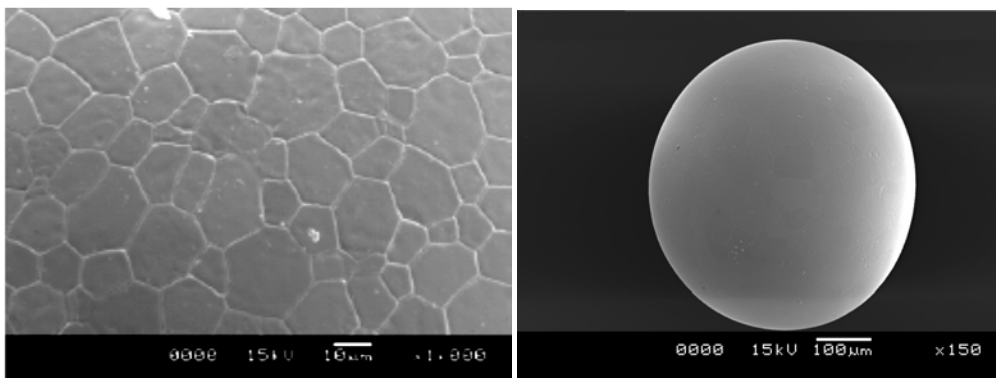


Figure 1 : SEM picture of the surface of an unirradiated UO₂ particle.

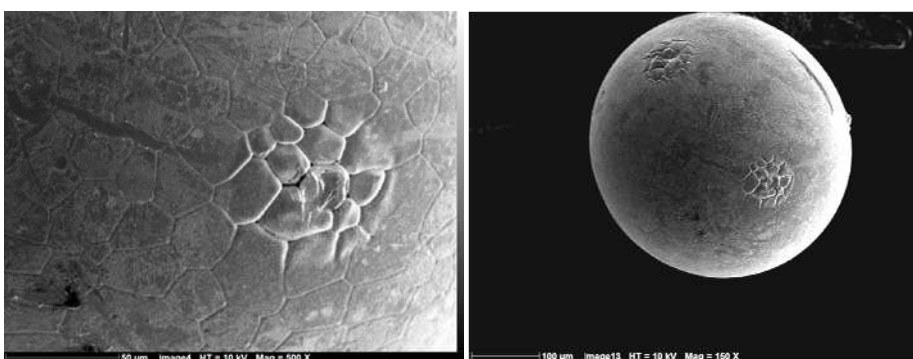


Figure 2 : SEM pictures of the surface of a UO₂ particle corroded in open to air atmosphere.

The spectrum of the unirradiated UO₂ is presented in **Figure 3**. It presents two characteristic bands around 445 cm⁻¹ and 560 cm⁻¹. The former is affected to the triply degenerated Raman active mode (T_{2g}) of U-O stretch (Amme et al., 2002; Amme et al., 2005; Desgranges et al., 2012; He and Shoesmith, 2010; Manara and Renker, 2003) whereas the latter characterizes defects present initially in the UO₂ fluorite structure (He and Shoesmith, 2010).

To estimate the Raman spectral modifications brought by the radiolytic products of water in the different studied atmospheres, a comparison was made between the UO₂ spectra before and after irradiation. The results showed that after irradiation in close atmospheres, the spectra of the UO₂ particles remained stable in function of the dose (not presented in this paper). When water in contact with the UO₂ particles was irradiated in open to air atmosphere, no spectral modifications were observed until 21.8 kGy deposited in the solution. From 43.7 kGy, modifications can be seen between 800 and 900 cm⁻¹ and appear clearly at the dose of 87.3 kGy. Two peaks appear indeed at 825 and 860 cm⁻¹. They characterize the metastudtite (UO₂(O₂) 2H₂O) (Amme et al., 2002; Canizarès et al., 2012; Guimbretière et al., 2011) formed on the UO₂ surface by oxidation with H₂O₂ produced by water radiolysis according to the reaction (i) (Corbel et al., 2006) :



After irradiation, temporal evolution of the solid particles was followed by Raman spectroscopy. The results showed that Raman spectra of the samples irradiated with an irradiation time $t = 1$ min remained unchanged after two months of the experiment. It seems that in these conditions, the concentration of H₂O₂ produced by water radiolysis was not sufficient to oxidize the UO₂ surface. For $t \geq 3$ min, Raman spectra of the irradiated particles evolved slightly in function of time until one week after irradiation and then remained stable.

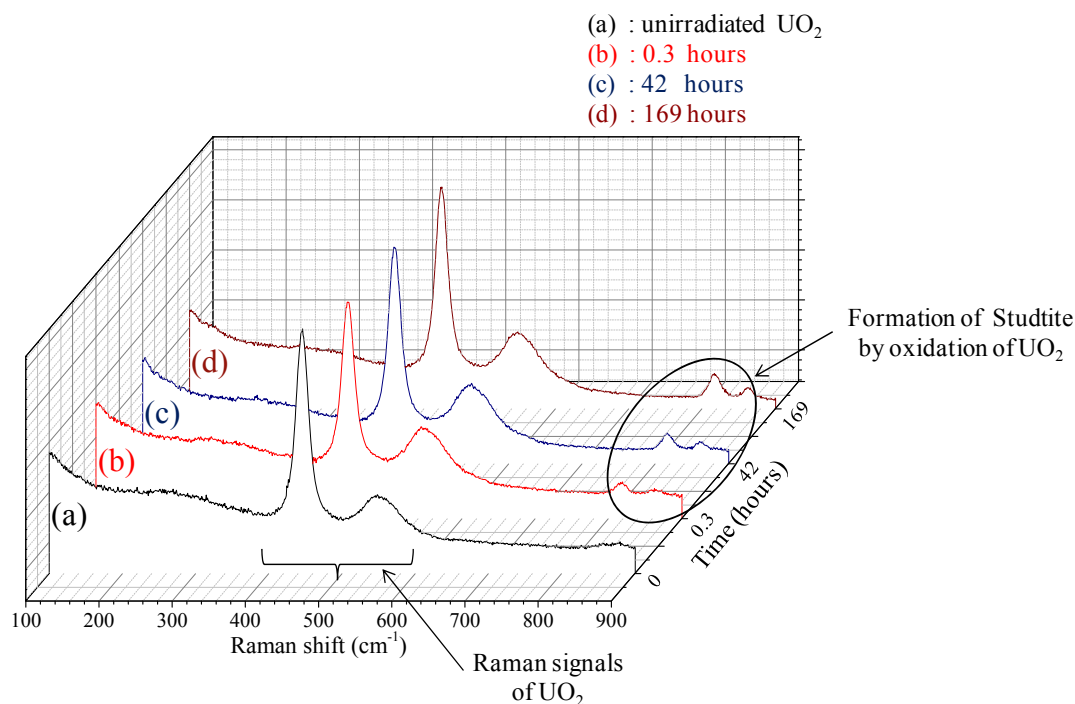


Figure 3 : Raman spectra of unirradiated UO₂ ((a)) and a sample irradiated at 8.73 kGy under open atmosphere ((b), (c) and (d)). The spectra (b), (c) and (d) are obtained 0.3, 42 and 169 hours after contact with the irradiated solution.

2. Uranium Solubility and Dissolution Rate

After irradiation, water in contact with the UO₂ particles was collected to quantify the soluble uranium species leached into the solution. Indeed, the metastudtite phase formed on the UO₂ surface after oxidation by H₂O₂ produced by water radiolysis underwent later a dissolution step where uranyl ions (UO₂²⁺) were leached into the solution. The mechanism of this dissolution reaction in ultra-pure water was rarely discussed in the literature. According to some authors (Satonnay et al., 2001), the dissolution step occurs according to reaction (ii) :



However, (Satonnay et al., 2001) proposed this dissolution mechanism based on a pH decreasing observed in their case after the solution irradiation. But, in our work measurement of the irradiated solutions acidity after each irradiation showed that pH did not evolve significantly and remained between 6.5 and 7.5. This means that the above proposed mechanism (reaction (ii)) does not occur in our case.

We have followed the variation of the soluble uranium species concentration leached into the solution after irradiation in function of the dose and the irradiation time in both open and close atmospheres. Regardless the irradiation atmosphere, the quantity of soluble uranium species increased with the dose and then remained constant from 43.6 kGy. At this latter dose, concentration of soluble uranium species found in close with air atmosphere ($5.25 \cdot 10^{-7} \text{ mol l}^{-1}$) is two-fold lower than that observed in open to air atmosphere ($11.49 \cdot 10^{-7} \text{ mol l}^{-1}$) due to inhibition of UO₂ oxidation by H₂ produced by water radiolysis. Values of [U] measured in our system at the steady state are presented in **Table 1** for an irradiation time of 10 min together with the normalized leaching rate (r). Moreover, **Table 1** presents also UO₂ dissolution rate values obtained from literature in order to compare it to our data. When UO₂ particles were irradiated between 10 and 20 min in close with air atmosphere, concentration of Uranium species remained practically constant. The obtained values were respectively 4.44 ± 0.5 and $5.25 \pm 0.5 \cdot 10^{-7} \text{ mol l}^{-1}$ for the latter mentioned irradiation time. These values are

comparable to most solubility data reported in the literature (Casas et al., 2009; Grambow et al., 2004; Jégou et al., 2003; Shoesmith and Sunder, 1992). However, we believe that in our work the thermodynamic equilibrium of the UO₂ particles dissolution was not achieved due to their composition of grains and grain boundaries. In these conditions, the dissolution rate was equal to 13.5 g m⁻² d⁻¹. A comparison of our dissolution rates with literature data is given in **Table 1**. Despite the similarity with [U] reported in the literature, r values obtained in this investigation are at least 10³ fold higher. It seems then that with the irradiation conditions used in our work, the UO₂ oxidation/dissolution mechanism occurred with a faster kinetic. This is very probably due to the impact of the high doses deposited into the irradiated solution for this work. In order to apprehend the dose impact on the UO₂ corrosion, we propose here a new methodology to express the dissolution rate. This methodology consists in calculating the dissolution rate in function of the dose deposited into the solution and not in function of the irradiation time. Thus, the new dissolution rate r' can be defined as follow :

$$r' = \frac{V}{S} \frac{d[U]}{dd} M_{UO_2}$$

Where $d[U]/dd$ (mol m⁻³) is the variation of soluble uranium species in function of the dose deposited into the solution. In our case the r' values are 2.01 µg m⁻² Gy⁻¹ for the open with air atmosphere experiment and 1.11 µg m⁻² Gy⁻¹ for the close with air atmosphere one. These results confirm that the reducing atmosphere (in presence of H₂) has a strong diminishing impact onto the UO₂ dissolution rate. Moreover, we have recalculated the same new dissolution rate r' from literature data (Grambow et al., 2004; Jégou et al., 2003) (**Table 1**) in order to compare it to our results. Even after taking into account the effect of the dose, r' values obtained in this investigation still higher (**table 1**). However, the dose rate used in our work is very high. It appears then that the dose rate is the main parameter which controls the UO₂ dissolution rate under irradiation.

Table 1 : Steady state U concentrations, Dissolution rates (r and r') of UO₂ and experimental conditions used in this study and others found in the literature.

| Ref. | Physical aspect UO ₂ | ⁴ He ²⁺ Beam | Dose Rate Gy min ⁻¹ | Atm. | [U] 10 ⁻⁷ mol l ⁻¹ | Diss. rate r mg m ⁻² d ⁻¹ | Diss. rate r' µg m ⁻² Gy ⁻¹ |
|------------------------------|---------------------------------|------------------------------------|--------------------------------|------|--|---|---|
| This work | Particle | External source* | 4366 | Ox | 11.49 | 13500 | 2.01 |
| | | | | Red | 5.25 | 7000 | 1.11 |
| (Shoesmith and Sunder, 1992) | Pellet | Internal | No | Ox | | 0.5 | X |
| (Casas et al., 2009) | Particle | No | 0 | Red | | 23.3 | X |
| (Grambow et al., 2004) | Colloid | External source* | 52.7 | Red | | 15.7 | 0.246 |
| (Jégou et al., 2003) | Pellet | Internal** | 18 | Ox | | 2.5 | 0.095 |
| | | | 1.8 | Ox | | 0.2 | 0.076 |

* ⁴He²⁺ Beam produced by cyclotron facility

** Doped UO₂ with Pu

3. H₂ and H₂O₂ Chemical Yields

When the UO₂ particles were irradiated in close with air atmosphere, H₂ produced by water radiolysis was quantified by calculating its radiolytic yield (G) in order to compare it to that of pure water and explain any gas consumption due to the inhibition of the UO₂ corrosion

process. In our work, the solution used during irradiation is ultra-pure water ($\rho = 1 \text{ kg l}^{-1}$) and the mean $G(\text{H}_2)$ was $0.02 \pm 0.002 \text{ } \mu\text{mol J}^{-1}$. It corresponds to the slope of the curve representing the variation of H_2 concentration in function of the dose. (Crumière et al., 2013) irradiated ultra-pure water by a cyclotron ${}^4\text{He}^{2+}$ beam with a 64.7 MeV energy and doses up to 800 Gy in Argon saturated atmosphere. $G(\text{H}_2)$ obtained in their work was $0.061 \pm 0.006 \text{ } \mu\text{mol J}^{-1}$ which is three-fold higher than that found in this investigation ($G(\text{H}_2) = 0.02 \pm 0.002 \text{ } \mu\text{mol J}^{-1}$). The irradiation conditions used by (Crumière et al., 2013) are very similar to those used in our work which means that for the system $\text{UO}_2/\text{ultra-pure water}$, a non negligible part (around 2/3) of the H_2 produced by water radiolysis was consumed certainly by inhibition of the UO_2 oxidation process.

In both open to air and close with air atmospheres, concentrations of H_2O_2 produced by water radiolysis was determined by the Ghormley method (Ghormley and Hochanadel, 1954) to analyze the effect of the irradiation atmosphere on its consumption by the UO_2 oxidation process. During this study, we have followed the variation of H_2O_2 concentration produced by water radiolysis under open to air and close with air atmospheres in function of the dose. It shows that concentration of H_2O_2 produced by water radiolysis increased with the dose and then reached a steady state from 21.8 kGy and 65.5 kGy in open to air and close with air atmospheres respectively. At these latter doses, $[\text{H}_2\text{O}_2]$ values were respectively $1.4 \cdot 10^{-4}$ and $5.5 \cdot 10^{-4} \text{ mol l}^{-1}$ which means that H_2O_2 concentration was four-fold higher when the UO_2 particles were irradiated in presence of H_2 . It seems then that in the latter conditions consumption of H_2O_2 and thus oxidation of the UO_2 surface was mainly inhibited by H_2 produced by water radiolysis. Radiolytic yields of H_2O_2 were also calculated and the values obtained were 0.06 ± 0.006 and $0.1 \pm 0.01 \text{ } \mu\text{mol J}^{-1}$ in open to air and close with air atmospheres respectively. $G(\text{H}_2\text{O}_2)$ obtained in open to air atmosphere is two fold lower than that obtained in close with air one. In this last condition, H_2O_2 yield is equal to that of ultra-pure water (LaVerne, 2004; Pastina and LaVerne, 1999; Yamashita et al., 2008a; Yamashita et al., 2008b) which means that this molecule was not consumed whereas a non-negligible part of H_2 was consumed. From this, we can conclude that in open to air atmosphere, half of the H_2O_2 produced by water radiolysis was consumed by oxidation of the UO_2 surface leading to the formation of metastudtite secondary phase. We can conclude too that H_2 has an inhibition effect on the UO_2 oxidation and this inhibition does not take place by direct effect on H_2O_2 but by interaction between H_2 and the UO_2 surface.

Conclusions and Future work

In conclusion, this work brings some light on the radiolytic corrosion of UO_2 by identification of (1) the oxidized secondary phase formed (metastudtite) and following its temporal evolution, (2) the H_2 role as an inhibitor agent, (3) the oxidative role of H_2O_2 and (4) the quantity of U species released. Detailed mechanisms of UO_2 corrosion/oxidation will be proposed taking into account the phenomena of water radiolysis.

Further studies will be performed in order to determine the role of the radical species produced by water radiolysis on UO_2 corrosion. Moreover, spacial resolution technique will be used in future surface characterization experiments in order to localize the corrosion on the grain boundaries as shown with SEM pictures.

Acknowledgement

We acknowledge B.Humbert and J.Y. Mevellec from the IMN laboratory for support in Raman measurements and N.Stephant for SEM measurements. The authors acknowledge ARRONAX staff for the efficient performing of irradiation runs onto the cyclotron facility. *The research leading to these results has received funding from the European Atomic Energy Community's (Euratom) Seventh Framework Programme FP7/2007-2011 under grant agreement n° 295722 (FIRST-Nuclides project).*

References

- Amme, M., Renker, B., Schmid, B., Feth, M.P., Bertagnolli, H., Döbelin, W., 2002. Raman microspectrometric identification of corrosion products formed on UO₂ nuclear fuel during leaching experiments. *Journal of Nuclear Materials* 306, 202-212.
- Amme, M., Wiss, T., Thiele, H., Boulet, P., Lang, H., 2005. Uranium secondary phase formation during anoxic hydrothermal leaching processes of UO₂ nuclear fuel. *Journal of Nuclear Materials* 341, 209-223.
- Canizarès, A., Guimbretière, G., Tobon, Y.A., Raimboux, N., Omnée, R., Perdicakis, M., Muzeau, B., Leoni, E., Alam, M.S., Mendes, E., Simon, D., Matzen, G., Corbel, C., Barthe, M.F., Simon, P., 2012. In situ Raman monitoring of materials under irradiation: study of uranium dioxide alteration by water radiolysis. *Journal of Raman Spectroscopy* 43, 1492-1497.
- Casas, I., de Pablo, J., Clarens, F., Giménez, J., Merino, J., Bruno, J., Martinez-Esparza, A., 2009. Combined effect of H₂O₂ and HCO₃⁻ on UO₂(s) dissolution rates under anoxic conditions. *Radiochimica Acta* 97, 485-490.
- Corbel, C., Sattonnay, G., Guilbert, S., Garrido, F., Barthe, M.F., Jegou, C., 2006. Addition versus radiolytic production effects of hydrogen peroxide on aqueous corrosion of UO₂. *Journal of Nuclear Materials* 348, 1-17.
- Crumièrè, F., Vandendorre, J., Essehli, R., Blain, G., Barbet, J., Fattahi, M., 2013. LET effects on the hydrogen production induced by the radiolysis of pure water. *Radiation Physics and Chemistry* 82, 74-79.
- Desgranges, L., Baldinozzi, G., Simon, P., Guimbretière, G., Canizares, A., 2012. Raman spectrum of U₄O₉: a new interpretation of damage lines in UO₂. *Journal of Raman Spectroscopy* 43, 455-458.
- Ghormley, J.A., Hochanadel, C.J., 1954. The Yields of Hydrogen and Hydrogen Peroxide in the irradiation of oxygen saturated water with Cobalt γ -Rays. *Journal of the American Chemical Society* 76, 3351-3352.
- Grambow, B., Mennecart, T., Fattahi, M., Blondiaux, G., 2004. Electrochemical aspects of radiolytically enhanced UO₂ dissolution. *Radiochimica Acta* 92, 603-609.
- Guimbretière, G., Canizares, A., Simon, P., Tobon-Correa, Y.A., Ammar, M.R., Corbel, C., Barthe, M.F., 2011. In-Situ Raman Observation of the First Step of Uranium Dioxide Weathering Exposed to Water Radiolysis. *Spectroscopy Letters* 44, 570-573.

He, H., Shoesmith, D., 2010. Raman spectroscopic studies of defect structures and phase transition in hyper-stoichiometric UO_{2+x} . *Physical Chemistry Chemical Physics* 12, 8108-8117.

Jégou, C., Broudic, V., Poulesquen, A., Bart, J.M., 2003. Effects of α and γ Radiolysis of Water on Alteration of the Spent UO_2 Nuclear Fuel Matrix. *MRS Online Proceedings Library* 807.

Kienzler, B., Metz, V., Duro, L., Valls, A., 2012. Collaborative Project "Fast / Instant Release of Safety Relevant Radionuclides from Spent Nuclear Fuel" , Budapest 09 - 11 October 2012 (KIT Scientific Reports ; 7639). *Proceedings 1st Annual Workshop (7th EC FP CP FIRST-Nuclides)*.

Manara, D., Renker, B., 2003. Raman spectra of stoichiometric and hyperstoichiometric uranium dioxide. *Journal of Nuclear Materials* 321, 233-237.

Sattonnay, G., Ardois, C., Corbel, C., Lucchini, J.F., Barthe, M.F., Garrido, F., Gosset, D., 2001. α -radiolysis effects on UO_2 alteration in water. *Journal of Nuclear Materials* 288, 11-19.

Shoesmith, D.W., Sunder, S., 1992. The prediction of nuclear fuel (UO_2) dissolution rates under waste disposal conditions. *Journal of Nuclear Materials* 190, 20-35.